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Abstract

Aim: Plant functional traits summarize the main variability in plant form and function across taxa and biomes. We assess whether geographic range size, climatic niche size, and local abundance of plants can be predicted by sets of traits (trait syndromes) or are driven by single traits.

Location: Eurasia.

Methods: Species distribution maps were extracted from the Chorological Database Halle to derive information on the geographic range size and climatic niche size for 456 herbaceous, dwarf shrub and shrub species. We estimated local species abundances based on 740,113 vegetation plots from the European Vegetation Archive, where abundances were available as plant species cover per plot. We compiled a complete species-by-trait matrix of 20 plant functional traits from trait databases (TRY, BioFlor and CLO-PLA). The relationships of species' geographic range size, climatic niche size and local abundance with single traits and trait syndromes were tested with multiple linear regression models.

Results: Generally, traits were more strongly related to local abundances than to broad-scale species distribution patterns in geographic and climatic space (range and niche size), but both were better predicted by trait combinations than by single traits. Local abundance increased with leaf area and specific leaf area (SLA). Geographic range size and climatic niche size both increased with SLA. While range size increased with plant height, niche size decreased with leaf carbon content.

Conclusion: Functional traits matter for species' abundance and distribution at both local and broad geographic scale. Local abundances are associated with different combinations of traits as compared to broad-scale distributions, pointing to filtering by different environmental and ecological factors acting at distinct spatial scales. However, traits related to the leaf economics spectrum were important for species' abundance and occurrence at both spatial scales. This finding emphasizes the general importance of resource acquisition strategies for the abundance and distribution of herbaceous, dwarf shrub and shrub species.

KEYWORDS

chorological database Halle (CDH), climatic niche, commonness and rarity, European Vegetation Archive (EVA), functional traits, geographic range, macroecology, vegetation-plot data

1 | INTRODUCTION

A central aim in functional macroecology is to understand to what degree plant morphological and physiological traits (Violle et al., 2007) affect species' distributions at large spatial scales (i.e. occurrences

across the geographic and climatic space), as well as local abundances within communities (Brown, 1995; McGill et al., 2006). The limits of species' broad-scale distributions reflect the interplay between dispersal barriers and bioclimatic conditions that affect range dynamics and evolution (Baselga et al., 2012). In contrast, species'

local abundance depends on factors operating at the local scale of species assemblages, such as habitat suitability, the local combination of environmental conditions, and biotic interactions (Peterson et al., 2011; Staniczenko et al., 2017). Under the assumption that species' functional traits reflect the mechanisms through which species respond to abiotic and biotic conditions to maximize their fitness, these traits are expected to predict both broad-scale distribution and local abundances (Suding et al., 2008; Heino & Tolonen, 2018).

Species can be rare or common (i.e. less or more abundant) within a local plant community. Similarly, some species have restricted distribution ranges while others are geographically widely distributed (Rabinowitz, 1981; Gurevitch et al., 2002; Enquist et al., 2019). It has been observed that species with larger geographic range sizes tend to have broader environmental tolerances (i.e. broader climatic niches), while geographically narrowly distributed species are also more likely to be narrowly distributed in climatic space (Slatyer et al., 2013; Sporbert et al., 2020). A positive relationship between climatic niche size and geographic range size across species thus seems to be a general macroecological pattern (Gaston, 2000; Slatyer et al., 2013; Cardillo et al., 2019). A species' local abundance results from population growth and demographical performance (Peterson et al., 2011). Within the geographic distribution range of a species, its local abundance at the community level is often highly variable. At the local scale, species abundance values are frequently used as descriptors of species performance and are an important characteristic of the composition of herbaceous plant communities (Kent and Coker, 1992; Chiarucci et al., 1999). In general, locally rare species tend to have a sparse cover in plant communities (Murray & Lepschi, 2004). Thus, potentially, local cover could also be considered a proxy for local rarity or commonness. However, local cover is in general low at most sites and high at only a few sites across a species' distribution range (Murphy et al., 2006). In contrast to

“everywhere sparse” species, these “somewhere abundant” species are reflected in right-skewed species abundance distributions, a common pattern in plant community ecology (McNellie et al., 2019). This skewness in local abundance might be caused by the distribution of optimal ecological conditions, and thus, might be causally linked to functional traits. As mean abundance across the species range itself does not capture the full variability of skewed frequency distributions, it should be considered together with the skewness of a species' cover value across its distribution range as proxies for rarity or commonness.

Functional traits have been used as proxies for species' dispersal abilities (Greene and Johnson 1993; Thompson et al., 2011), environmental tolerances (Loehle, 1998; Bohner & Diez, 2020) or competitiveness (Kunstler et al., 2016). Specific functional traits have been linked to commonness and rarity on both local and large scales (see Table 1). For example, studies have found plant height, used as a proxy for competitive ability, to be positively correlated with range size, with taller species more widespread than shorter ones (Lavergne et al., 2004; Kolb et al., 2006). Similarly, on the local scale, common (i.e. more abundant) species have been associated with taller stature and with other traits that are proxies for species' physiological activity and productivity, including larger specific leaf area (SLA) and higher leaf nitrogen (N) content (Grime et al., 1997; Hegde & Ellstrand, 1999; Lavergne et al., 2004; Mariotte, 2014; Lachaise et al., 2020). Nitrogen (N) and phosphorus (P) availabilities limit plant growth in most terrestrial ecosystems (Güsewell, 2004). Low nutrient availability (e.g. phosphorus limitation) may weaken the relationship between productivity-related traits and macroclimate (Bruehlheide et al., 2018). As a consequence, there might be a negative correlation between species' N:P ratio and both their local abundance and broad-scale distribution. Regarding species' persistence, locally more abundant species have been associated with perennial life cycle and clonal growth (Eriksson & Jakobsson, 1998; Kolb et al., 2006). In contrast, at large spatial scales, rare

TABLE 1 Traits used in this study, their function in the community, and their reported correlation with local abundance and broad-scale distribution being unimodal (—), positive (↑) or negative (↓)

Trait	Function	Reported correlation with	
		Local abundance	Broad-scale distribution
Specific leaf area, leaf C, leaf N, leaf P, leaf dry matter content	Productivity, competitive ability, leaf economics spectrum	Specific leaf area ↑ ¹ , — ³ , leaf N ↑ ¹	Specific leaf area — ² , ↑ ³ , leaf N — ² , leaf dry matter content — ²
Leaf N:P ratio	Nutrient supply	↓ ^{4,15}	↓ ^{4,15}
Plant height, leaf area	Competitive ability	Plant height ↑ ^{1,2,5,6} , leaf area ↑ ¹	Plant height — ⁷ , ↑ ^{2,8}
Seed mass, seed number per reproductive unit, dispersal unit length	Dispersal, regeneration	Seed mass — ^{7,8,14} , ↓ ¹⁴ , seed number per reproductive unit ↓ ^{4,8}	Seed mass — ^{2,10} , ↑ ⁸ , ↓ ^{11,12} , seed number per reproductive unit ↑ ^{2,8,9}
Life cycle, life form, clonality	Persistence	Perennials ↑ ^{7,8} , clonal growth ↑ ^{7,8}	Annuals — ⁸ , perennials — ⁸ , therophytes ↑ ⁸ , phanerophytes ↓ ⁸ , clonal growth — ⁷ , ↓ ⁷

¹Mariotte (2014); ²Lavergne et al. (2004); ³Lachaise et al. (2020); ⁴Bruehlheide et al. (2018); ⁵Hedge & Ellstrand (1999); ⁶Grime et al. (1997); ⁷Eriksson & Jakobsson (1998); ⁸Kolb et al. (2006); ⁹Van der Veken et al. (2007); ¹⁰Thompson et al. (1999); ¹¹Guo et al. (2000); ¹²Oakwood et al. (1993); ¹³Kelly & Woodward (1996); ¹⁴Leishman & Murray (2001); ¹⁵Güsewell (2004).

species have been associated with prevailing clonal growth (Kelly & Woodward, 1996) and woodiness (Oakwood et al., 1993). Several studies have investigated the relationships linking dispersal or regeneration-related traits with species' local abundance and broad-scale distribution patterns. On the local scale, more abundant species were found to produce fewer and lighter seeds than rare species (Hedge & Ellstrand, 1999; Guo et al., 2000; Kolb et al., 2006). In contrast, at large spatial scales, geographically widespread species have been found to produce significantly more and heavier seeds than small-ranged plant species (Lavergne et al., 2004; Kolb et al., 2006; Van der Veken et al., 2007).

While some studies have found relationships between functional traits and local abundance and/or broad-scale distribution patterns, others have failed to detect a clear correlation (see Table 1). So far, the majority of studies have focused on single traits rather than on trait combinations or trait syndromes (but see Díaz et al., 2016; Guo et al., 2018) as predictors of large and local distribution patterns. However, no single trait can completely describe a species' ecological strategy (Winemiller et al., 2015; Marino et al., 2020). Rather, species' local abundance and broad-scale distribution patterns might be affected by different sets of traits (Marino et al., 2020). It has been suggested that locally rare and geographically restricted plant species differ systematically from more common species in functional traits that are related to species' productivity, competitive ability, dispersal, regeneration and persistence (Murray et al., 2002). However, the different states and values of traits cannot be unconditionally combined. Díaz et al. (2004) highlighted that the functional space occupied by vascular plant species is strongly constrained by trade-offs between traits. On the one hand, the leaf economics spectrum describes a productivity–persistence trade-off and contrasts species with a set of successful trait combinations for quick returns on investments of nutrients and dry mass in leaves to species with a slower potential rate of return of more persistent leaves (Wright et al., 2004). On the other hand, the size spectrum reflects the species' life cycle, with small stature species, smaller seeds and short lifespans vs long-lived woody plants (Díaz et al., 2016; Table 1).

In this study, we aimed at unravelling the relationships between traits (single traits or trait syndromes) and species distributions at broad spatial scale and abundances at local scale. Specifically, we focused on 20 traits that are expected to respond to bioclimatic drivers and capture the essence of plant life forms and functions (Wright et al., 2004; Petchey & Gaston, 2006; Díaz et al., 2016; Bruelheide et al., 2018). We tested for these relationships across 456 European herbaceous, dwarf shrub and shrub species by interrelating existing data on functional traits with the species' (a) geographic range size; (b) climatic niche size; and (c) local abundance, which was measured as (i) mean cover from all the vegetation plots in which a species was present and (ii) skewness of cover values. We expected climatic niche size and geographic range size to be driven by the same underlying environmental factors and ecological processes (Colwell & Rangel, 2009), and therefore to be positively correlated, and to be predicted by many of the same single traits or trait syndromes (Table 1). We aimed to answer

the following research questions: (a) can single plant functional traits or sets of traits (trait syndromes) best explain the local abundance (i.e. a measure of commonness) and broad-scale distributions of plant species; and (b) do the specific traits and their relative contributions to species' abundance and distribution differ between the local and broad spatial scales?

2 | METHODS

2.1 | Broad-scale distribution metrics: geographic range size and climatic niche size

We used available digitized species distribution data (i.e. range polygons and point occurrences) of the Chorological Database Halle (CDH; E. Welk et al., unpublished data) to assess the Eurasian geographic ranges of 456 herbaceous, dwarf shrub and shrub species, including their neophytic occurrences. A list of these species can be found in Appendix S1. In total, CDH stores information on species distribution ranges for more than 17,000 vascular plant species but expert-drawn range maps were compiled for 5,583 taxa based on national and floristic databases and maps from the floristic literature (Tralau, 1969-1981; Lundquist & Nordenstam, 1988; Lundquist, 1992; Lundquist & Jäger, 1995-2007). These data are published as distribution range maps (Meusel et al., 1965, 1978; Meusel & Jäger, 1992). We used the subset of these species that met the criteria described below. Data stored in CDH can be requested for research objectives via chorologie.biologie.uni-halle.de/choro/.

We aggregated species' point and polygon distribution data using a raster grid layer of 2.5 arc-min resolution, which corresponds to grid cells covering approximately 15 km² each across Central Europe. As a measure of range size for each species, we counted the number of grid cells occupied (approximating the area of occupancy in the geographical space).

We determined the multi-dimensional climatic space (or climatic niche) of each geographic range based on principal component analysis (PCA) of 19 bioclimatic variables from the WorldClim 2.0 database (Fick & Hijmans, 2017), also at 2.5 arc-min resolution. The resulting global background climatic space was well represented by the first two principal components, which accounted for 70.75% of the total climatic variance. The two-dimensional PCA space was rasterized into 100 × 100 PCA grid cells, considered as the background climatic niche, as explained in Appendix S2. The species' niche size was then calculated as the number of PCA grid cells occupied in the climatic space (i.e., the area of occupancy in the bioclimatic niche space; for detailed information see Appendix S2).

2.2 | Local abundance metrics in vegetation plots: mean cover and skewness of cover values

As a measure of local abundance, percentage cover values were obtained for each of the study species in 740,113 vegetation plots from

the European Vegetation Archive (EVA; Chytrý et al., 2016), queried in October 2015. Overall, EVA comprised vegetation plots from all European countries plus Turkey, Georgia, Armenia, Azerbaijan and parts of Russia. We included vegetation plots from all vegetation types available from EVA, except aquatic vegetation. We matched synonymous species names according to the taxonomic reference list for Germany (German SL version 1.2, Jansen & Dengler, 2008) and to four taxonomic reference lists available via the R package *taxize* (Chamberlain & Szöcs, 2013; R Core Team, 2018), i.e. Encyclopedia of Life (EOL), International Plant Names Index (IPNI), Integrated Taxonomic Information Service (ITIS) and Tropicos. In cases where no exact match was found, taxon names were resolved using the Taxonomic Name Resolution Service (TNRS) and all names matched or converted from a synonym were considered accepted taxon names when probabilities were $\geq 95\%$. We merged the data for subspecies at the species level following the taxonomic hierarchy in TNRS. The selected study species occurred within at least 100 vegetation plots in the EVA dataset. Vegetation plots with a geographic location uncertainty of more than 10 km were removed prior to this selection. The median occurrence (i.e. number of vegetation

plots a species occurred in) per species was 2,162 (interquartile range 846 to 5,137). Information on source databases that provided vegetation-plot data can be found in Appendix S3. Cover or cover abundance values that were based on ordinal scales (e.g. Domin, 1928; Braun-Blanquet, 1951) were converted to percentage cover (van der Maarel, 1979).

For each species, we measured two aspects of the species' abundance across the vegetation plots. First, we calculated its "mean cover": the arithmetic mean of the percentage cover values from all the vegetation plots at 2.5 arc-min raster cells in which the species was present in EVA. Second, we evaluated the frequency distribution of these percentage cover values (see Figure 1 for details on the procedure for three example species). For this, we computed the shape of the distribution function of the percentage cover values. In general, those values are not normally, exponentially or log-normally distributed (Figure 1a–c); thus, we developed a non-parametric approach for measuring the shape of the distribution function. This was achieved by calculating the distribution quantiles in 5% steps, resulting in 20 quantile values. We then fitted a non-linear model on those 20 quantile values and obtained the estimate and the credible

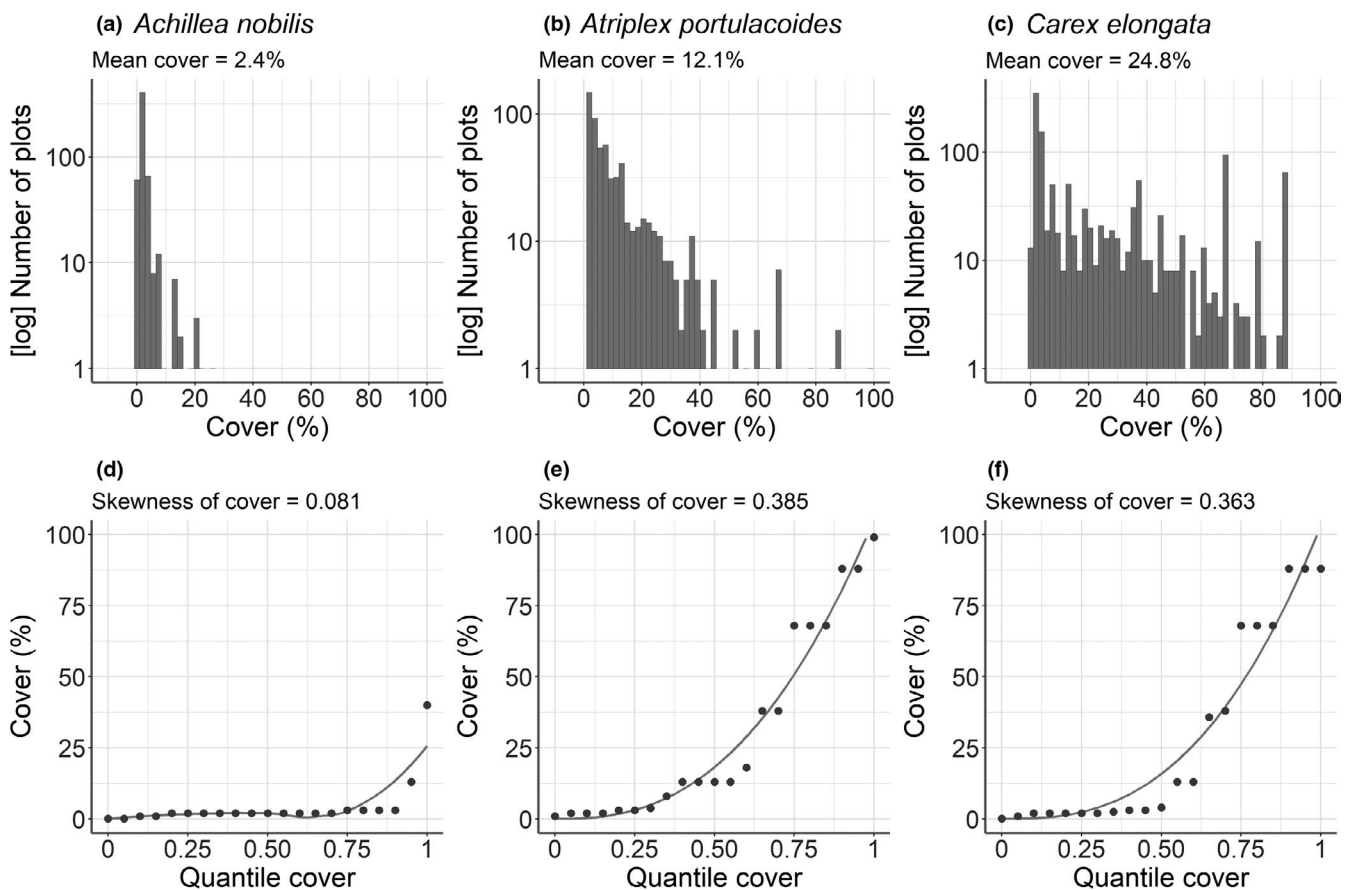


FIGURE 1 Examples of distribution of species' cover values from vegetation plots and calculated mean cover value for (a) *Achillea nobilis*, the species with the lowest mean cover value, (b) *Atriplex portulacoides*, a species with intermediate mean cover value and (c) *Carex elongata*, the species with the highest mean cover value. Note the log scale for frequency. Distribution quantiles from species' cover values were calculated and used to compute the shape of the frequency distribution function for each species, respectively (d–f). Non-linear models on the extracted quantile values were applied to calculate the area under the histograms of cover values (AUH), ranging from 0 to 1, with values close to 0 indicating a strongly right-skewed distribution whereas values close to 1 point to a strongly left-skewed distribution of cover values

interval of the area under the histogram (AUH) (Figure 1d–f). We applied a Bayesian Markov chain Monte Carlo (MCMC) method following Feng et al. (2017), using an exponential distribution, 0.95 confidence level and 10,000 iterations. The resulting AUH value for a given species ranged from 0 to 1, with values lower or higher than 0.5 meaning that the distribution of cover values for a focal species is right- or left-skewed, respectively. The lower the AUH value, the higher was the rarity (i.e. the proportion of relatively low cover values). Thus, the AUH values are suitable as proxies for abundance structure across the vegetation plots. Hereafter, we refer to the AUH values as “skewness of cover values” and use it as an alternative metric, additional to mean cover, to assess across-plot species abundance.

2.3 | Explanatory variables: plant functional traits

We compiled a complete species trait matrix with 20 plant functional traits (see Table 2 and Appendix S1). The trait matrix included nine binary variables: five for life form following Raunkiaer (1934); three for life cycle (derived from BiolFlor database; Kühn et al., 2004); and one for clonality (derived from the CLO-PLA database;

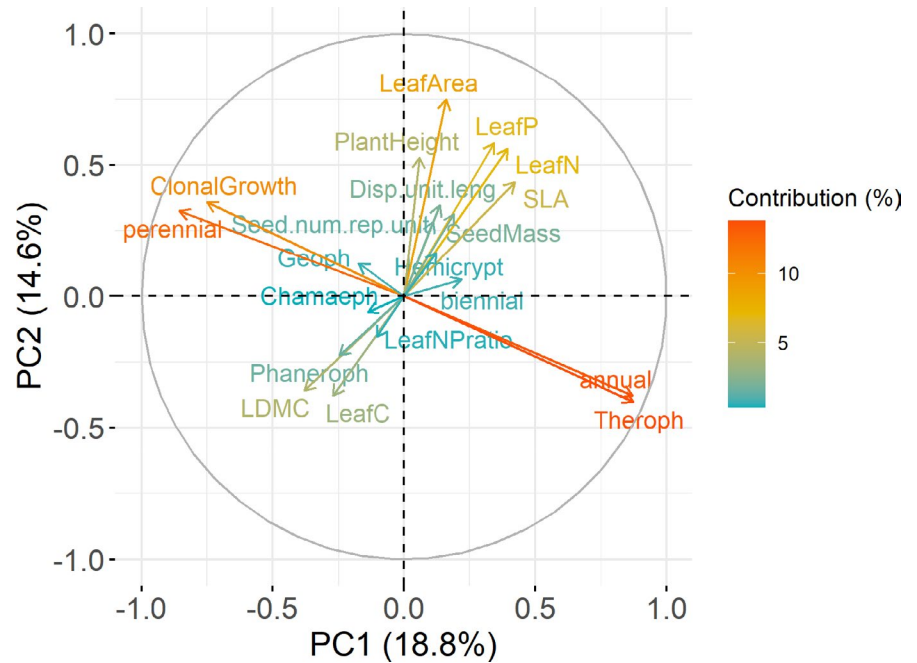
Klimešová et al., 2017). We included information on 11 continuous trait variables from the global plant-trait database TRY (Kattge et al., 2020). All continuous TRY trait values were derived from Bruelheide et al. (2018) who applied a gap-filling approach with Bayesian Hierarchical Probabilistic Matrix Factorization (BHPMF; Schrodt et al., 2015) to fill gaps in the observed species-by-trait matrix data received from TRY. Continuous trait variables were ln-transformed prior to analysis.

A PCA of the 20 traits included in this study was generated using the package *factoextra* (Kassambara & Mundt, 2017), allowing the visualization of the trait contributions (loadings) to the first and second principal components (Figure 2). The first component corresponded to traits of life form (i.e. therophyte), life (i.e. annual and perennial) and clonal growth and accounted for 18.8% of the total variation in trait values. The second component, corresponding to leaf traits, accounted for 14.6% of the total variation in trait values. The third and fourth components corresponded to dispersal traits (i.e. seed mass and dispersal unit length) and life form (i.e. hemicryptophyte) and accounted for 11.4% and 8.7% of the total variation in trait values, respectively. The mean and standard deviation of observed trait values are given in Table 2. A correlation matrix including the correlation coefficients of all pairwise trait combinations can be

Trait	Abbreviation	Unit	Mean	SD
Leaf area	LeafArea	mm ²	2,128.74	6,346.69
Specific leaf area	SLA	m ² /kg	23.30	8.99
Leaf C content	LeafC	mg/g	451.36	24.93
Leaf N content	LeafN	mg/g	24.59	7.61
Leaf P content	LeafP	mg/g	2.14	0.95
Leaf dry matter content	LDMC	mg/g	0.22	0.08
Leaf N:P ratio	LeafNPratio	g/g	12.13	6.35
Plant height	PlantHeight	m	0.41	0.44
Seed mass	SeedMass	mg	2.36	4.59
Seed number per reproductive unit	SeedNumRepUnit		42,956.49	447,429.20
Dispersal unit length	DispUnitLeng	mm	3.18	2.29
Life cycle length				
Annual	Annual	Proportion [%]	16.2	-
Biennial	Biennial	Proportion [%]	9.4	-
Perennial	Perennial	Proportion [%]	81.8	-
Life form				-
Phanerophyte	Phaneroph	Proportion [%]	5.0	-
Chamaephyte	Chamaeph	Proportion [%]	6.1	-
Hemicryptophyte	Hemicrypt	Proportion [%]	78.2	-
Geophyte	Geoph	Proportion [%]	10.7	-
Therophyte	Theroph	Proportion [%]	14.3	-
Clonality				-
Clonal growth	ClonalGrowth	Proportion [%]	88.8	-

TABLE 2 Traits, abbreviations of trait names, units of measurement, and mean and standard deviation of observed trait values

FIGURE 2 Principal component analysis of the 20 traits included in this study. Colour represents the trait contributions (%) to the PCA (first and second components). The first and second components accounted for 18.8% and 14.6% of the total variation in trait values, respectively. For abbreviation of the trait names see Table 2



found in Appendix S4. The values of trait contributions (loadings) to all PCA axes and the explained variation in trait values are also given in Appendix S4.

2.4 | Statistical modelling: linking plant functional traits to mean cover values, skewness of cover values, geographic range size, and climatic niche size

We used the function `phylo.maker` from the package *V. PhyloMaker* (Jin & Qian, 2019) to create a phylogenetic tree of the studied species. The function `phylo4d` from the package *phylobase* (Hackathon et al., 2013) was applied to link trait data to the species' phylogeny. We applied Pagel's Lambda statistic (Pagel, 1999) and Fritz and Purvis' *D* (Fritz & Purvis, 2010) statistics to quantify the strength of phylogenetic signal among the 456 studied species for each of the 20 studied trait variables. Pagel's Lambda statistic revealed a strong phylogenetic signal in all continuous trait variables. Fritz and Purvis' *D* revealed a phylogenetic signal in all binary trait variables (see Appendix S4). Therefore, we ran phylogenetic generalized least-squares models (i.e. a phylogenetically corrected model) using the function `pgls` from the package *caper* (Orme et al., 2018), that takes into account the phylogenetic non-independence between species trait values when analysing the relationships linking plant functional traits to mean cover values, skewness of cover values, geographic range size, and climatic niche size. We also applied phylogenetic generalized least-squares models to examine the relationship between species' geographic range size and climatic niche size, as well as between species' mean cover values and skewness of cover values. To allow fair comparisons of the magnitude of effect sizes between continuous and binary trait variables, all continuous and ln-transformed trait variables were standardized by subtracting the mean and dividing by two standard deviations (Gelman, 2008). The lasso

procedure in function `glmnet` from the package *glmnet* (Friedman et al., 2010) was applied to extract those of the 20 trait variables that contributed the most in the four linear models (mean cover values, skewness of cover values, geographic range size, and climatic niche size being the four studied response variables). The minimum lambda ratio value in each of the four linear models was 0.005, 0.0005, 0.0001 and 0.01 for mean cover values, skewness of cover values, geographic range size and climatic niche size, respectively (see Appendix S4 for the predictor variables included in the models). To test the effect of trait combinations and interactions for each of the four response variables, we then tested all possible combinations of the predictor variables that contributed the most, including their two-way interaction terms, applying the dredge function from the *MuMIn* package (Barton, 2019). We allowed for a maximum of three predictor terms to be included in a given candidate model ($m.max = 3$); with this, univariate models were applied for single traits and multivariate models for trait combinations (see Appendix S4 for the lists of candidate models for the four response variables). Finally, the Akaike Information Criterion (AIC), with $\Delta AIC < 2$ was used to identify the most parsimonious candidate model with a maximum of three predictor terms for each of the four studied response variables. We computed the variance inflation factor (VIF) for each predictor term in the most parsimonious models to check for potential multicollinearity issues among the continuous predictor variables, using the function `vif` from the package *car* (Fox & Weisberg, 2019).

As a complementary analysis, we ran another set of phylogenetic generalized least-squares models but using principal components (PCs) from the PCA on the trait space as explanatory variables instead of using the original set of trait variables. Because PCs represent uncorrelated dimensions of trait values, this analysis provides an alternative approach to test for trait syndromes on species' local abundance and distribution patterns. We used all the species scores on all 20 PCs as predictor terms in the models described above and tested for

combinations and interactions between PCs in the same way as described for traits. By applying the dredge function we tested all possible combinations of the predictor variables that contributed the most, including two-way interaction between PCs, for each of the four response variables. AIC with $\Delta\text{AIC} < 2$ was used to identify the most parsimonious candidate model with a maximum of three predictor terms for each of the four studied response variables (see Appendix S4 for the trait contributions [loadings] to all 20 PCs).

3 | RESULTS

3.1 | Broad-scale distribution metrics: geographic range size and climatic niche size

Species' range size (number of occupied grid cells in geographical space) ranged from 1,947 in *Dactylorhiza sambucina* to 782,025 in *Stellaria media* with a median range size of 310,070 cells. Species' climatic niche size (number of occupied PCA grid cells within the climatic niche space) ranged from 162 in *Scabiosa canescens* to 9,318 in *Plantago major* with a median of 3,236 cells. We found a positive relationship between species' geographic range and climatic niche size ($R^2 = 0.605$, p -value < 0.001 in a phylogenetically corrected model; Appendix S4).

3.2 | Local abundance metrics: mean cover and skewness of cover values

Species' mean cover from all the vegetation plots in which a species was present ranged from 2.4% for *Achillea nobilis* to 24.8% for *Carex*

elongata (Figure 1a and c). The interquartile range of was 4.6% to 8.1% and the median was 5.9%. Species' skewness of cover values ranged from 0.081 (strongly right-skewed distribution of low cover values) in *Achillea nobilis* to 0.385 in *Atriplex portulacoides* (Figure 1d and e). The interquartile range was 0.158–0.226 and the median was 0.180. Species' mean cover was positively related to species' skewness of cover values in a phylogenetically corrected model ($R^2 = 0.763$, p -value < 0.001 ; Appendix S4).

3.3 | The contribution of functional traits to explaining values and skewness of cover values

All response variables were better explained by trait combinations than by single traits, e.g. among the list of candidate models for mean cover as response variable, the best univariate model with plant height as predictor variable was ranked 17 (see Appendix S4). SLA was the strongest predictor for all response variables; with species with high SLA having larger range sizes, broader climatic niche sizes, and higher local abundances. For each specific response variable, SLA interacted with other traits to give scale-specific and different trait responses. Specifically, geographic range size was larger in species with taller stature and lower leaf N:P ratio. In contrast, climatic niche size was larger in species that had lower leaf C content. The mentioned functional traits were significantly, though not strongly related to species' geographic range size ($R^2 = 0.090$, p -value < 0.001) and climatic niche size ($R^2 = 0.069$, p -value < 0.001) in the phylogenetic generalized least-squares models (Table 3, Figure 3a and b). Species' mean cover and the skewness of cover values was higher in species with higher SLA value and with higher leaf area values. The interaction of the variables SLA and leaf area was

Response variable	Multiple R^2 ; p -value	Predictor terms entered in model	Regression coefficient	p -value	VIF
Geographic range size	0.090; < 0.001	SLA (m^2/kg)	0.213	< 0.001	1.015
		Plant height (m)	0.140	< 0.01	1.014
		Leaf N:P ratio (g/g)	-0.140	< 0.01	1.020
Climatic niche size	0.069; < 0.001	SLA (m^2/kg)	0.181	< 0.001	1.063
		Leaf C (mg/g)	-0.126	< 0.01	1.043
		Therophyte	0.108	0.096	1.024
Mean cover	0.211; < 0.001	SLA (m^2/kg)	2.146	< 0.001	1.066
		Leaf area (mm^2)	1.417	< 0.001	1.076
		Leaf area (mm^2) * SLA (m^2/kg)	1.722	< 0.001	1.041
Skewness of cover values	0.169; < 0.001	SLA (m^2/kg)	0.033	< 0.001	1.066
		Leaf area (mm^2)	0.025	< 0.001	1.076
		Leaf area (mm^2) * SLA (m^2/kg)	0.030	< 0.001	1.041

TABLE 3 Effects of species traits on species geographic range size, climatic niche size, mean cover and skewness of cover values

Note: Akaike Information Criterion (AIC) was used to identify the most parsimonious model for each of the four response variables. Interaction terms are indicated by *. Computed variance inflation factor (VIF) for each predictor variable was low, indicating that the predictors were not correlated with each other.

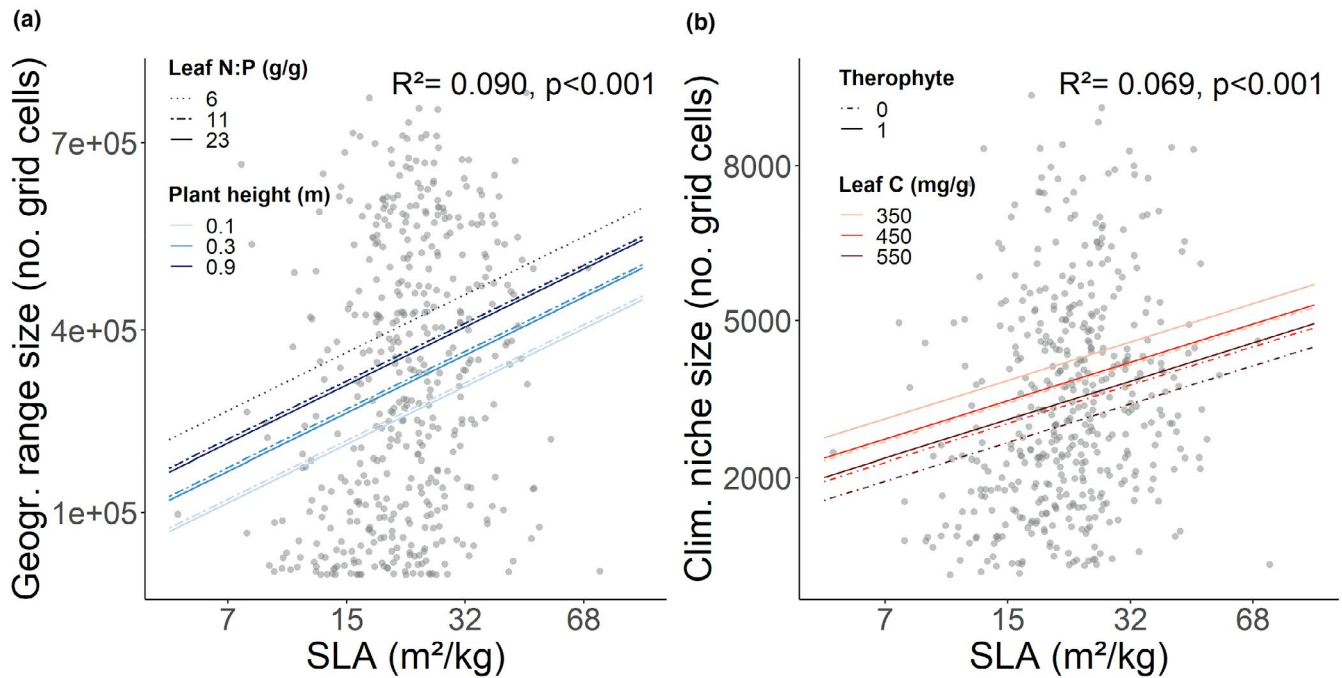


FIGURE 3 Scatter plot of observed values and regression lines from phylogenetic generalized least-squares models, showing the effects of the three most predictive terms on species' (a) geographic range size and (b) climatic niche size. For geographic range size, coloured and dashed lines represent the 5th, 50th and 95th percentile in values for plant height and leaf N:P ratio, respectively; for climatic niche size, coloured lines, solid and dotted, represent the 5th, 50th and 95th percentile in leaf C content values for therophytic and non-therophytic species, respectively

positive and these functional traits were significantly related to both species mean cover ($R^2 = 0.211$, p -value < 0.001) and the AUH measure of the skewness of cover values ($R^2 = 0.169$, p -value < 0.001 ; Table 3, Figure 4a and b).

The specific traits identified in the final multivariate models for the four response variables had also high loadings on the PCs that were identified as important predictors in the multivariate PCA-based models (see Appendix S4 for the trait contributions [loadings] for all 20 PCs). In addition, the axes captured some more traits with maximum absolute loadings that were not selected in the final trait-based models, such as leaf area for geographic range size, leaf P content for climatic niche size, clonal growth for mean cover and leaf area for the skewness of cover values. However, for each of the four response variables, the three PCs in the final models explained less variation than the trait-based models: geographic range size ($R^2 = 0.063$, p -value < 0.001 ; in the sequence of importance, the model included PCs 12, 2 and 1), climatic niche size ($R^2 = 0.069$, p -value < 0.001 ; based on PCs 1, 12 and 19), mean cover ($R^2 = 0.123$, p -value < 0.001 ; based on PCs 13, 2 and 4) and skewness of cover values ($R^2 = 0.094$, p -value < 0.001 ; based on PCs 6, 2 and 4).

4 | DISCUSSION

Species' local abundances (i.e. a measure of commonness) were more strongly related to traits than were species' broad-scale distribution patterns in the geographic and climatic space. This indicates

that plant traits better capture local processes acting at the community level (such as biotic processes) than broad-scale macroecological processes. Both local abundances and broad-scale distribution patterns were better predicted by combinations of traits than by single traits.

Specific leaf area had a significant positive effect and explained most of the observed variation in all four models predicting species' local abundance and broad-scale distribution patterns. SLA is a productivity- and competitive ability-related trait, that reflects species strategies for rapid acquisition of resources, with higher SLA values allowing a species to capture more light for a given biomass investment in leaves, for example (Díaz et al., 2004; Wright et al., 2004; Mariotte, 2014). In line with our findings, several studies state common species to be associated with higher SLA (Grime et al., 1997; Díaz et al., 2004; Mariotte, 2014; Lachaise et al., 2020). While species' local abundances were best predicted by the interaction between leaf area and SLA, reflecting the leaf economics spectrum trait syndrome (Díaz et al., 2004), broad-scale distribution metrics were best predicted by different combinations of traits. While geographic range size increased with increasing plant height, climatic niche size decreased with increasing leaf carbon content, and both increased with increasing SLA.

At the local scale, leaf area showed a significantly positive effect on species abundance. This result offers a functional explanation that species with larger leaves, allowing better light capture, are able to attain higher local abundances than species with smaller leaves (Mariotte, 2014). Moreover, leaf area was particularly important in

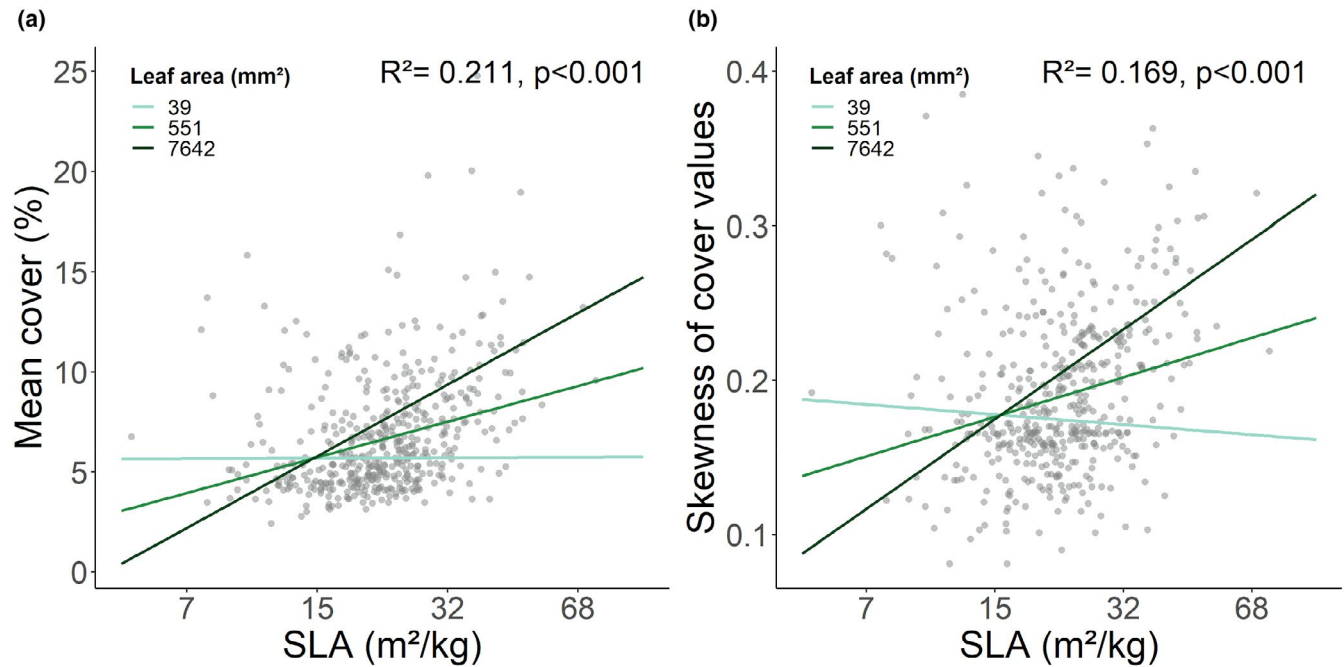


FIGURE 4 Scatter plot of observed values and regression lines from phylogenetic generalized least-squares models, showing the effects of the relevant species traits on species' local abundance. Plots show the effect of the interaction between leaf area and SLA on (c) mean cover and (d) skewness of cover values. Coloured lines represent the 5th, 50th and 95th percentile in values for leaf area

interaction with SLA values, as species' local abundance was higher in species with large leaves and high SLA values. Like SLA, leaf area is interpreted as a trait that is positively related to species' productivity and competitive ability (Diaz et al., 2004; Wright et al., 2004). This indicates that being nearer the "fast" end of the leaf economics spectrum (i.e. trait syndrome) tends to increase local abundance. In contrast, the ability to grow clonally was not selected in any of the final models based on original trait variables. The ability of species for clonal growth plays an important role both in short-distance spread and in persistence within habitats (Benot et al., 2013) and previous studies found clonality to be positively associated with local abundance (Eriksson & Jakobsson, 1998; Kolb et al., 2006). Accordingly, this trait had high axis loadings in the PC-based model with mean cover value as response variable. The absence of clonal growth from our final models based on original trait values is probably due to its negative correlation with SLA in our set of species.

At large spatial extent, geographic range size of species was positively related to plant height (i.e. with taller species being more widespread). High stature is known to have a competitive advantage and to be associated with common species. Greater plant height of widely distributed species suggests that these species may have higher competitive ability for space and light than narrowly distributed species. A similar positive correlation between plant height and geographic range size was found for herbaceous species in the French Mediterranean region (Lavergne et al., 2004) and in temperate forests in Germany (Kolb et al., 2006). We found leaf N:P ratio to be negatively correlated with geographic range size. Nitrogen (N) and phosphorus (P) availability can limit plant growth in terrestrial ecosystems, and N:P ratios are on average higher in stress-tolerant species compared to ruderals

(Güsewell, 2004). Ruderal species are characterized by rapid growth and they establish much quicker and thrive better in disturbed habitats than stress-tolerant and competitor species (Grime, 1979; Wright et al., 2004; Guo et al., 2018) and generally undergo long-distance dispersal (Baker, 1965). Thus, a plausible explanation is that ruderality has a positive effect on species' geographic range size. As shown in a global study by Bruelheide et al. (2018), species' leaf N:P ratio declines at higher latitudes. Many species primarily found in boreal regions obtain broader geographic range sizes in comparison to species mainly found further south (e.g. in Mediterranean regions), presumably because of post-glacial re-expansion. This might be another plausible explanation for the negative relationship between geographic range size and leaf N:P ratio in our study.

In our study, the distribution range in climatic space was larger in species with lower leaf carbon content, even when accounting for SLA. In general, carbon content is expected to be negatively related to SLA (Reich, 2014), but both traits seem to explain independent variation in climatic niche size. This was brought about by species with broad climatic niche sizes, for which SLA alone was a poor predictor, such as species with a tendency to succulence (e.g. *Plantago major*), which have leaves with low SLA but yet low leaf carbon content. This indicates that species following a "fast" strategy, according to the leaf economics spectrum, are better adapted to obtain broad climatic niche sizes. Species with a therophytic life form (i.e. annual plants that overwinter as a seed) did show marginally greater climatic tolerance (i.e. broader climatic niche size). A short generation time is a selective advantage to annuals over biennials and perennials (Pysek & Richardson, 2007), with annual species being capable of faster reproduction and spread by seeds than perennials, e.g. many weed species are annual ruderal

species that generally undergo long-distance dispersal (Baker, 1965). Finally, traits related to dispersal, regeneration and persistence were not significantly correlated with local abundance or broad-scale distribution in our models. These results confirm previous studies that found no significant relationship between species' abundance or broad-scale distribution and seed mass (Thompson et al., 1999; Leishman & Murray, 2001; Lavergne et al., 2004) or life cycle length (Kolb et al., 2006).

Our results largely confirm trends previously reported about the existing association between species' geographic range and climatic niche size, with widely distributed species also having broad climatic tolerances and geographically narrowly distributed species also narrowly distributed in climatic space (Gaston, 2000; Slatyer et al., 2013; Cardillo et al., 2019). We found an overall right-skewed distribution in cover values for most of the studied species, with species exhibiting low cover at most sites and high cover in only a few sites across their distribution range. The species mean cover values were positively related to the skewness of cover values. Therefore, for our species set, we consider the measure of skewness, calculated as the AUH, a robust tool to capture both the mean and the variability of cover values across a species' whole distribution range.

Nevertheless, single plant traits and trait syndromes only weakly explained the total observed variation in species' broad-scale distribution metrics. We see two plausible explanations for the weak predictive power of functional traits on species' broad-scale distribution metrics.

First, our study was carried out on the species' whole Eurasian distribution range, which includes a wide range of habitat types and bioclimatic zones. Species functional traits are expected to be related to those environmental conditions under which the species occurs (Lavorel & Garnier, 2002; McGill et al., 2006). With this, both widely distributed and geographically restricted species might be characterized by the same traits in different habitats, vegetation types or geographic regions, depending on the local conditions (Aerts & Chapin, 2000). Therefore, future studies should incorporate habitat variability, by means of comparisons among single habitat types (e.g. by applying EUNIS habitat classification; Chytrý et al., 2020), and test for consistency of the role of traits for patterns of species commonness at different spatial scales.

Second, in this study, we used mean trait values derived from trait databases. Several studies have provided evidence that functional traits express not only species-specific characteristics, but also intraspecific variability in leaf traits (Reich & Oleksyn, 2004; Albert et al., 2011; Moles et al., 2014; Niinemets, 2015; Wright et al., 2017). This intraspecific trait variation may influence the interactions among and between species and their environment and, therefore, might influence species performance (Bolnick et al., 2003; Siefert et al., 2015). Therefore, we encourage future studies to include intraspecific trait variation in addition to mean values for species traits when investigating studies over large geographic scales.

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AUTHOR CONTRIBUTIONS

MS, EW and HB conceived the study. GS and MS harmonized data retrieved from EVA and CDH. MS harmonized data retrieved from TRY with data retrieved from EVA and CDH. EW, GS and MS developed the measures for niche properties and abundance skewness. MS and HB carried out statistical analyses. MS produced the graphs. MS and HB wrote the paper, MS led the writing. All other authors contributed data, discussed the results and commented on, and/or substantially edited, the manuscript.

DATA AVAILABILITY STATEMENT

The data used in this paper are from large multi-contributor databases (EVA, TRY). They cannot be made publicly available because of the third-party ownership issues. The data selections released for this study are stored in internal repositories of the source databases and can be made available for re-analyses upon request. The EVA dataset is stored in the EVA repository with reference to project no. 24.

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REFERENCES

- Aerts, R. & Chapin, F.S. III (2000) The mineral nutrition of wild plants revisited: a re-evaluation of processes and patterns. *Advances in Ecological Research*, 30, 1–67. [https://doi.org/10.1016/S0065-2504\(08\)60016-1](https://doi.org/10.1016/S0065-2504(08)60016-1)
- Albert, C.H., Grassein, F., Schurr, F.M., Vieilledent, G. & Violle, C. (2011) When and how should intraspecific variability be considered in trait-based plant ecology? *Perspectives in Plant Ecology, Evolution and Systematics*, 13, 217–225. <https://doi.org/10.1016/j.ppees.2011.04.003>
- Baker, H.G. (1965) Characteristics and modes of origin of weeds. In: Baker, H.G. & Stebbins, G.L. (Eds.) *The genetics of colonizing species*. Academic Press, pp. 147–172.
- Barton, K. (2019) *MuMIn: Multi-Model Inference. R package version 1.43.15*. Available at <https://CRAN.R-project.org/package=MuMIn> [Accessed 17 January 2021]
- Baselga, A., Lobo, J.M., Svenning, J.C. & Araújo, M.B. (2012) Global patterns in the shape of species geographical ranges reveal range determinants. *Journal of Biogeography*, 39, 760–771. <https://doi.org/10.1111/j.1365-2699.2011.02612.x>
- Benot, M.L., Bittebiere, A.K., Ernoult, A., Clement, B. & Mony, C. (2013) Fine-scale spatial patterns in grassland communities depend on species clonal dispersal ability and interactions with neighbours. *Journal of Ecology*, 101, 626–636. <https://doi.org/10.1111/1365-2745.12066>
- Bohner, T. & Diez, J. (2020) Extensive mismatches between species distributions and performance and their relationship to functional traits. *Ecology Letters*, 23, 33–44. <https://doi.org/10.1111/ele.13396>
- Bolnick, D.I., Svanbäck, R., Fordyce, J.A., Yang, L.H., Davis, J.M., Hulsey, C.D. et al (2003) The ecology of individuals: Incidence and implications of individual specialization. *The American Naturalist*, 161, 1–28.
- Braun-Blanquet, J. (1951) *Pflanzensoziologie: Grundzüge der Vegetationskunde*, 2nd edition. Springer.
- Brown, J.H. (1995) *Macroecology*. University of Chicago Press.
- Bruelheide, H., Dengler, J., Purschke, O., Lenoir, J., Jiménez-Alfaro, B., Hennekens, S.M. et al (2018) Global trait–environment relationships of plant communities. *Nature Ecology & Evolution*, 2, 1906–1917. <https://doi.org/10.1038/s41559-018-0699-8>
- Cardillo, M., Dinnage, R. & McAlister, W. (2019) The relationship between environmental niche breadth and geographic range size across plant species. *Journal of Biogeography*, 46, 97–109. <https://doi.org/10.1111/jbi.13477>
- Chamberlain, S.A. & Szöcs, E. (2013) taxize: taxonomic search and retrieval in R. *F1000Research*, 2, 191. <https://doi.org/10.12688/f1000research.2-191.v2>
- Chiarucci, A., Wilson, J.B., Anderson, B.J. & De Dominicis, V. (1999) Cover versus biomass as an estimate of species abundance: does it make a difference to the conclusions? *Journal of Vegetation Science*, 10, 35–42. <https://doi.org/10.2307/3237158>
- Chytrý, M., Hennekens, S.M., Jiménez-Alfaro, B., Knollová, I., Dengler, J., Jansen, F. et al (2016) European Vegetation Archive (EVA): An integrated database of European vegetation plots. *Applied Vegetation Science*, 19, 173–180. <https://doi.org/10.1111/avsc.12191>
- Chytrý, M., Tichý, L., Hennekens, S.M., Knollová, I., Janssen, J.A., Rodwell, J.S. et al (2020) EUNIS Habitat Classification: expert system, characteristic species combinations and distribution maps of European habitats. *Applied Vegetation Science*, 23, 648–675. <https://doi.org/10.1111/avsc.12519>
- Colwell, R.K. & Rangel, T.F. (2009) Hutchinson's duality: The once and future niche. *Proceedings of the National Academy of Sciences*, 106, 19651–19658. <https://doi.org/10.1073/pnas.0901650106>
- Díaz, S., Hodgson, J.G., Thompson, K., Cabido, M., Cornelissen, J.H., Jalili, A. et al (2004) The plant traits that drive ecosystems: evidence from three continents. *Journal of Vegetation Science*, 15, 295–304. <https://doi.org/10.1111/j.1654-1103.2004.tb02266.x>
- Díaz, S., Kattge, J., Cornelissen, J.H., Wright, I.J., Lavorel, S., Dray, S. et al (2016) The global spectrum of plant form and function. *Nature*, 529, 167. <https://doi.org/10.1038/nature16489>
- Domin, K. (1928) The relations of the Tatra mountain vegetation to the edaphic factors of the habitat: A synecological study. *Acta Botanica Bohemica*, 6, 133–163.
- Enquist, B.J., Feng, X., Boyle, B., Maitner, B., Newman, E.A., Jørgensen, P.M. et al (2019) The commonness of rarity: Global and future distribution of rarity across land plants. *Science Advances*, 5, eaaz0414. <https://doi.org/10.1126/sciadv.aaz0414>
- Eriksson, O. & Jakobsson, A. (1998) Abundance, distribution and life histories of grassland plants: a comparative study of 81 species. *Journal of Ecology*, 86, 922–933. <https://doi.org/10.1046/j.1365-2745.1998.00309.x>
- Feng, D., Cortese, G. & Baumgartner, R. (2017) A comparison of confidence/credible interval methods for the area under the ROC curve for continuous diagnostic tests with small sample size. *Statistical Methods in Medical Research*, 26, 2603–2621. <https://doi.org/10.1177/0962280215602040>
- Fick, S.E. & Hijmans, R.J. (2017) WorldClim 2: new 1-km spatial resolution climate surfaces for global land areas. *International Journal of Climatology*, 37, 4302–4315. <https://doi.org/10.1002/joc.5086>
- Fox, J. & Weisberg, S. (2019) *An R Companion to Applied Regression*. 3rd Edition. Sage.
- Friedman, J., Hastie, T. & Tibshirani, R. (2010) Regularization paths for generalized linear models via coordinate descent. *Journal of Statistical Software*, 33, 1–22.
- Fritz, S.A. & Purvis, A. (2010) Selectivity in mammalian extinction risk and threat types: a new measure of phylogenetic signal strength in binary traits. *Conservation Biology*, 24, 1042–1051. <https://doi.org/10.1111/j.1523-1739.2010.01455.x>
- Gaston, K.J. (2000) Global patterns in biodiversity. *Nature*, 405, 220–227. <https://doi.org/10.1038/35012228>
- Gaston, K. & Blackburn, T. (2008) *Pattern and Process in Macroecology*. John Wiley & Sons.
- Gelman, A. (2008) Scaling regression inputs by dividing by two standard deviations. *Statistics in Medicine*, 27, 2865–2873. <https://doi.org/10.1002/sim.3107>
- Greene, D.F. & Johnson, E.A. (1993) Seed mass and dispersal capacity in wind-dispersed diaspores. *Oikos*, 67, 69–74. <https://doi.org/10.2307/3545096>
- Grime, J.P. (1979) *Plant Strategies and Vegetation Processes*. Wiley.
- Grime, J.P., Thompson, K., Hunt, R., Hodgson, J.G., Cornelissen, J.H.C., Rorison, I.H. et al (1997) Integrated screening validates primary axes of specialisation in plants. *Oikos*, 79, 259–281. <https://doi.org/10.2307/3546011>
- Guo, Q., Brown, J.H., Valone, T.J. & Kachman, S.D. (2000) Constraints of seed size on plant distribution and abundance. *Ecology*, 81, 2149–2155. [https://doi.org/10.1890/0012-9658\(2000\)081\[2149:COSSOP\]2.0.CO;2](https://doi.org/10.1890/0012-9658(2000)081[2149:COSSOP]2.0.CO;2)
- Guo, W.Y., van Kleunen, M., Winter, M., Weigelt, P., Stein, A., Pierce, S. et al (2018) The role of adaptive strategies in plant naturalization. *Ecology Letters*, 21, 1380–1389. <https://doi.org/10.1111/ele.13104>
- Gurevitch, J., Scheiner, S. & Fox, G. (2002) *The Ecology of Plants*. Sinauer Associates.
- Güsewell, S. (2004) N:P ratios in terrestrial plants: variation and functional significance. *New Phytologist*, 164, 243–266. <https://doi.org/10.1111/j.1469-8137.2004.01192.x>
- Hackathon, R., Bolker, B., Butler, M., Cowan, P., De Vienne, D., Eddelbuettel, D. et al (2013) *phylobase: Base package for phylogenetic structures and comparative data*. R package version 0.8, 4. Available at <https://cran.r-project.org/web/packages/phylobase/index.html> [Accessed 17 January 2021].
- Hegde, S.G. & Ellstrand, N.C. (1999) Life history differences between rare and common flowering plant species of California and the British



- Isles. *International Journal of Plant Sciences*, 160, 1083–1091. <https://doi.org/10.1086/314204>
- Heino, J. & Tolonen, K.T. (2018) Ecological niche features override biological traits and taxonomic relatedness as predictors of occupancy and abundance in lake littoral macroinvertebrates. *Ecography*, 41, 2092–2103. <https://doi.org/10.1111/ecog.03968>
- Jansen, F. & Dengler, J. (2008) GermanSL – eine universelle taxonomische Referenzliste für Vegetationsdatenbanken. *Tuexenia*, 28, 239–253.
- Jin, Y. & Qian, H. (2019) V. PhyloMaker: an R package that can generate very large phylogenies for vascular plants. *Ecography*, 42, 1353–1359. Available at <https://github.com/jinyizju/V.PhyloMaker> [Accessed 17 January 2021]
- Kassambara, A. & Mundt, F. (2017) *factoextra: Extract and Visualize the Results of Multivariate Data Analyses*. Available at <https://cran.r-project.org/web/packages/factoextra/index.html> Accessed 17 January 2021
- Kattge, J., Bönisch, G., Díaz, S., Lavorel, S., Prentice, I.C., Leadley, P. et al (2020) TRY plant trait database – enhanced coverage and open access. *Global Change Biology*, 26, 119–188. <https://doi.org/10.1111/gcb.14904>
- Kent, M., & Coker, P. (1992). *Vegetation description and analysis, a practical approach*. New York, NY: John Wiley & Sons.
- Kelly, C.K. & Woodward, F.I. (1996) Ecological correlates of plant range size: taxonomies and phylogenies in the study of plant commonness and rarity in Great Britain. *Philosophical Transactions of the Royal Society of London, Series B: Biological Sciences*, 351, 1261–1269. <https://doi.org/10.1098/rstb.1996.0109>
- Klímešová, J., Danihelka, J., Chrtěk, J., de Bello, F. & Herben, T. (2017) CLO-PLA: A database of clonal and bud-bank traits of the Central European flora. *Ecology*, 98, 1179. <https://doi.org/10.1002/ecy.1745>
- Kolb, A., Barsch, F. & Diekmann, M. (2006) Determinants of local abundance and range size in forest vascular plants. *Global Ecology and Biogeography*, 15, 237–247. <https://doi.org/10.1111/j.1466-8238.2005.00210.x>
- Kühn, I., Durka, W. & Klotz, S. (2004) BioFlor: a new plant-trait database as a tool for plant invasion ecology. *Diversity and Distributions*, 10, 363–365. Available at www.jstor.org/stable/3246738 [Accessed 17 January 2021]
- Kunstler, G., Falster, D., Coomes, D.A., Hui, F., Kooyman, R.M., Laughlin, D.C. et al (2016) Plant functional traits have globally consistent effects on competition. *Nature*, 529, 204–207. <https://doi.org/10.1038/nature16476>
- Lachaise, T., Bergmann, J., Rillig, M.C. & van Kleunen, M. (2020) Below- and aboveground traits explain local abundance, and regional, continental and global occurrence frequencies of grassland plants. *Oikos*, 130, 110–120. <https://doi.org/10.1111/oik.07874>
- Lavergne, S., Thompson, J.D., Garnier, E. & Debussche, M. (2004) The biology and ecology of narrow endemic and widespread plants: a comparative study of trait variation in 20 congeneric pairs. *Oikos*, 107, 505–518. <https://doi.org/10.1111/j.0030-1299.2004.13423.x>
- Lavorel, S. & Garnier, E. (2002) Predicting changes in community composition and ecosystem functioning from plant traits: revisiting the Holy Grail. *Functional Ecology*, 16, 545–556. <https://doi.org/10.1046/j.1365-2435.2002.00664.x>
- Leishman, M.R. & Murray, B.R. (2001) The relationship between seed size and abundance in plant communities: model predictions and observed patterns. *Oikos*, 94, 151–161. <https://doi.org/10.1034/j.1600-0706.2001.10392.x>
- Loehle, C. (1998) Height growth rate tradeoffs determine northern and southern range limits for trees. *Journal of Biogeography*, 25, 735–742. <https://doi.org/10.1046/j.1365-2699.1998.2540735.x>
- Lundqvist, J. (1992) *Index Holmiensis*. 7, Swedish Museum of Natural History.
- Lundqvist, J. & Jäger, E.J. (1995–2007) *Index Holmiensis*. Swedish Museum of Natural History.
- Lundqvist, J. & Nordenstam, B. (1988) *Index Holmiensis*. 6, Swedish Museum of Natural History.
- Marino, N.A., Céréghino, R., Gilbert, B., Petermann, J.S., Srivastava, D.S., de Omena, P.M. et al (2020) Species niches, not traits, determine abundance and occupancy patterns: A multi-site synthesis. *Global Ecology and Biogeography*, 29, 295–308. <https://doi.org/10.1111/geb.13029>
- Mariotte, P. (2014) Do subordinate species punch above their weight? Evidence from above- and below-ground. *New Phytologist*, 203, 16–21. <https://doi.org/10.1111/nph.12789>
- McGill, B.J., Enquist, B.J., Weiher, E. & Westoby, M. (2006) Rebuilding community ecology from functional traits. *Trends in Ecology & Evolution*, 21, 178–185. <https://doi.org/10.1016/j.tree.2006.02.002>
- McNellie, M.J., Dorrrough, J. & Oliver, I. (2019) Species abundance distributions should underpin ordinal cover abundance transformations. *Applied Vegetation Science*, 22, 361–372. <https://doi.org/10.1111/avsc.12437>
- Meusel, H., Jäger, E.J. & Weinert, E. (1965) *Vergleichende Chorologie der zentral-europäischen Flora, Karten*. I, VEB Gustav Fischer Verlag.
- Meusel, H., Jäger, E.J., Rauschert, S. & Weinert, E. (1978) *Vergleichende Chorologie der zentral-europäischen Flora, Karten*. II, VEB Gustav Fischer Verlag.
- Meusel, H. & Jäger, E.J. (1992) *Vergleichende Chorologie der zentral-europäischen Flora, Karten*. III, Gustav Fischer Verlag.
- Moles, A.T., Perkins, S.E., Laffan, S.W., Flores-Moreno, H., Awasthy, M., Tindall, M.L. et al (2014) Which is a better predictor of plant traits: temperature or precipitation? *Journal of Vegetation Science*, 25, 1167–1180. <https://doi.org/10.1111/jvs.12190>
- Murphy, H.T., VanDerWal, J. & Lovett-Doust, J. (2006) Distribution of abundance across the range in eastern North American trees. *Global Ecology and Biogeography*, 15, 63–71. <https://doi.org/10.1111/j.1466-822X.2006.00194.x>
- Murray, B.R. & Lepschi, B.J. (2004) Are locally rare species abundant elsewhere in their geographical range? *Austral Ecology*, 29, 287–293. <https://doi.org/10.1111/j.1442-9993.2004.01365.x>
- Murray, B.R., Thrall, P.H., Gill, A.M. & Nicotra, A.B. (2002) How plant life-history and ecological traits relate to species rarity and commonness at varying spatial scales. *Austral Ecology*, 27, 291–310. <https://doi.org/10.1046/j.1442-9993.2002.01181.x>
- Niinemets, Ü. (2015) Is there a species spectrum within the world-wide leaf economics spectrum? Major variations in leaf functional traits in the Mediterranean sclerophyll *Quercus ilex*. *New Phytologist*, 205, 79–96. <https://doi.org/10.1111/nph.13001>
- Oakwood, M., Jurado, E., Leishman, M. & Westoby, M. (1993) Geographic ranges of plant species in relation to dispersal morphology, growth form and diaspore weight. *Journal of Biogeography*, 20, 563–571. Available at www.jstor.org/stable/2845727 [Accessed 17 January 2021]
- Orme, D., Freckleton, R., Thomas, G., Petzoldt, T., Fritz, S., Isaac, N. & et al (2018) *caper: Comparative Analyses of Phylogenetics and Evolution in R. R package version 1.0.1*. Available at <https://CRAN.R-project.org/package=caper> [Accessed 17 January 2021]
- Pagel, M. (1999) Inferring the historical patterns of biological evolution. *Nature*, 401, 877–884. <https://doi.org/10.1038/44766>
- Petchey, O.L. & Gaston, K.J. (2006) Functional diversity: back to basics and looking forward. *Ecology Letters*, 9, 741–758. <https://doi.org/10.1111/j.1461-0248.2006.00924.x>
- Peterson, A.T., Soberón, J., Pearson, R.G., Anderson, R.P., Martínez-Meyer, E., Nakamura, M. & et al (2011) *Ecological niches and geographic distributions (MPB-49)*. 56, Princeton University Press.
- Pyšek, P. & Richardson, D.M. (2007) Traits associated with invasiveness in alien plants: where do we stand? In: Nentwig, W. (Ed.) *Biological Invasions*. Springer, pp. 97–125.
- R Core Team (2018) *R: A language and environment for statistical computing*. R Foundation for Statistical Computing.

- Rabinowitz, D. (1981) Seven forms of rarity. In: Synges, H. & Chichester, J. (Eds.) *The Biological Aspects of Rare Plant Conservation*. John Wiley & Sons, pp. 207–217.
- Raunkiaer, C. (1934) *The life forms of plants and statistical plant geography, being the collected papers of C. Raunkiaer*. Clarendon Press.
- Reich, P.B. (2014) The world-wide 'fast-slow' plant economics spectrum: a traits manifesto. *Journal of Ecology*, 102, 275–301. <https://doi.org/10.1111/1365-2745.12211>
- Reich, P.B. & Oleksyn, J. (2004) Global patterns of plant leaf N and P in relation to temperature and latitude. *Proceedings of the National Academy of Sciences*, 101, 11001–11006. <https://doi.org/10.1073/pnas.0403588101>
- Schrodt, F., Kattge, J., Shan, H., Fazayeli, F., Joswig, J., Banerjee, A. et al (2015) BHPMF – a hierarchical Bayesian approach to gap-filling and trait prediction for macroecology and functional biogeography. *Global Ecology and Biogeography*, 24, 1510–1521. <https://doi.org/10.1111/geb.12335>
- Siefert, A., Violle, C., Chalmandrier, L., Albert, C.H., Taudiere, A., Fajardo, A. et al (2015) A global meta-analysis of the relative extent of intraspecific trait variation in plant communities. *Ecology Letters*, 18, 1406–1419. <https://doi.org/10.1111/ele.12508>
- Slatyer, R.A., Hirst, M. & Sexton, J.P. (2013) Niche breadth predicts geographical range size: A general ecological pattern. *Ecology Letters*, 16, 1104–1114. <https://doi.org/10.1111/ele.12140>
- Sporbert, M., Keil, P., Seidler, G., Bruelheide, H., Jandt, U., Aćić, S. et al (2020) Testing macroecological abundance patterns: The relationship between local abundance and range size, range position and climatic suitability among European vascular plants. *Journal of Biogeography*, 47, 2210–2222. <https://doi.org/10.1111/jbi.13926>
- Staniczenko, P.P., Sivasubramaniam, P., Suttle, K.B. & Pearson, R.G. (2017) Linking macroecology and community ecology: refining predictions of species distributions using biotic interaction networks. *Ecology Letters*, 20, 693–707. <https://doi.org/10.1111/ele.12770>
- Suding, K.N., Lavorel, S., Chapin, F.S. III, Cornelissen, J.H., Díaz, S., Garnier, E. et al (2008) Scaling environmental change through the community-level: a trait-based response-and-effect framework for plants. *Global Change Biology*, 14, 1125–1140. <https://doi.org/10.1111/j.1365-2486.2008.01557.x>
- Thompson, F.J., Moles, A.T., Auld, T.D. & Kingsford, R.T. (2011) Seed dispersal distance is more strongly correlated with plant height than with seed mass. *Journal of Ecology*, 99, 1299–1307. <https://doi.org/10.1111/j.1365-2745.2011.01867.x>
- Thompson, K., Gaston, K.J. & Band, S.R. (1999) Range size, dispersal and niche breadth in the herbaceous flora of central England. *Journal of Ecology*, 87, 150–155. <https://doi.org/10.1046/j.1365-2745.1999.00334.x>
- Tralau, H. (1969–1981) *Index Holmiensis*. 1–5, Swedish Museum of Natural History.
- van der Maarel, E. (1979) Transformation of cover-abundance values in phytosociology and its effects on community similarity. *Vegetatio*, 39, 97–114.
- Van der Veken, S., Bellemare, J., Verheyen, K. & Hermy, M. (2007) Life-history traits are correlated with geographical distribution patterns of western European forest herb species. *Journal of Biogeography*, 34, 1723–1735. <https://doi.org/10.1111/j.1365-2699.2007.01738.x>
- Violle, C., Navas, M.L., Vile, D., Kazakou, E., Fortunel, C., Hummel, I. et al (2007) Let the concept of trait be functional. *Oikos*, 116, 882–892. <https://doi.org/10.1111/j.0030-1299.2007.15559.x>
- Winemiller, K.O., Fitzgerald, D.B., Bower, L.M. & Pianka, E.R. (2015) Functional traits, convergent evolution, and periodic tables of niches. *Ecology Letters*, 18, 737–751. <https://doi.org/10.1111/ele.12462>
- Wright, I.J., Dong, N., Maire, V., Prentice, I.C., Westoby, M., Díaz, S. et al (2017) Global climatic drivers of leaf size. *Science*, 357, 917–921. <https://doi.org/10.1126/science.aal4760>
- Wright, I.J., Reich, P.B., Westoby, M., Ackerly, D.D., Baruch, Z., Bongers, F. et al (2004) The worldwide leaf economics spectrum. *Nature*, 428, 821–827. <https://doi.org/10.1038/nature02403>

SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

Appendix S1. Species name; number of vegetation plots occupied per species; range and niche size; calculated mean abundance and skewness of cover values; trait values

Appendix S2. Climatic resampling procedure and background principal component analysis (PCA) niche space of the study area

Appendix S3. Information on the 59 databases that provided vegetation plots analysed in this paper

Appendix S4. Supporting information to the results

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Different sets of traits explain abundance and distribution patterns of European plants at different spatial scales

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